

# Scenario Simulation of Ecosystem Services Value Changes: Implication for Sustainability Development Goals

Shishay kiros weldegebriel<sup>1\*</sup>, Ranavijai Singh<sup>1</sup>, Kumelachew Yeshitela<sup>2</sup>

<sup>1</sup> Ethiopian Civil Service University, Ethiopia

<sup>2</sup> Ethiopian Institute of Architecture, Building Construction and City Development, Addis Ababa University, Ethiopia

**Abstract:-** This scientific investigation quantified future land use/ land cover variations on ecosystem service values under different scenarios in Mekelle city region for the year 2030. The study utilized land use/land cover dynamics data for the year 2019 and predicted LULC changes for the year 2030. Identification of the driving forces was the most important step in predicting the future LULC and was performed using the cellular automata models in CLUDMondo which is a future land use simulation model by coupling human and the natural driving forces. Scenarios were made under business as usual, implementation of PES schemes and spatial planning. In a business-as-usual scenario, it was detected that ecosystem services value will fall by 2030 from the base year of 2019, while under spatial planning scenario, the ecosystem service value will increase by 5.2%. The results can serve as valuable information on policy design in developing land use options, which aid to enhance ecosystem service value. Therefore, special attention should be given to the rehabilitation of ecosystems, the protection of the remaining natural vegetation and water bodies. Generally, these results offer useful inputs to planners to adopt eco-friendly policies to achieve sustainable development goals.

**Keywords:-** Scenario, Ecosystem Service, Value, Land Use/Land Change, Sustainable Development Goals.

## I. INTRODUCTION

Modelling for ecosystem services and scenario study are being more and more used to support decision making in land use(Ardavan Zarandiana, 2017). Prediction of futures landscape variations entails understanding and incorporation of historical landscape trends, the existing land change processes and feedbacks, and the incorporation of scenarios. Noticeably, monitoring and modelling of landscape dynamics strongly depend on the scale and aims of the planned applications(Houet, Verburg, & Loveland, 2010). Modelling ecosystem services (ES) is a crucial tool for the development of strategies that will guarantee future supply, provision and quantification(Olosutean, 2015).

The prediction of land-use alteration can improve the far-sightedness of land use planning, which has great implication for regional sustainable development. (Hu et al., 2020). The relationship between ES and land-use changes highlights the importance of ES in guiding land-use planning and ecosystem management to promote sustainability(Z. Li, Cheng, & Han, 2020). The scenario modelling of ecosystem services can provide the scientific basis for sustainable ecosystem management. (Fu et al., 2018). Monitoring the impact of current Land Use/Land Cover (LULC) management practices on future Ecosystem Services (ESs) provisioning has been emphasized because of the effect of such practices on ecological sustainability(Yirsaw, Wu, Shi, Temesgen, & Bekele, 2017).

Ecosystem services valuation is now commonly accredited for its optimistic role in achieving the three pillars of sustainable development(Pandeya et al., 2016). Mekelle's Urbanisation is an inevitable tendency, and the economic development that accompanies urbanisation can have negative implications and influences for ecosystem services. However, Mekelle city can play a crucial part in the solution if nature is considered an important parts of the city urban planning. Mekelle city tends to expand, in areas important for biodiversity such as watersheds. Biodiversity and ecosystem services both within cities and beyond the boundaries are important for the urban population as they contribute to food and drinking water, regulate temperature, absorb pollution, reduce vulnerabilities and disaster risks, and offer liveable places and recreational opportunities that contribute to human well-being.

Unplanned urbanization is likely to result in ecosystem loss when natural habitats are fragmented through city extensions. Despite their importance, these ecosystem services in the study area are poorly understood, scarcely monitored at the city-region level, and as a result, there is rapid degradation and depletion. There is greater uncertainty about future ecosystem services supply in Mekelle city region. Monitoring the negative consequences of LULC while sustaining the supply of essential ecosystem services is vital. Therefore, this study intended to quantify the land-use changes over time and predict the future ecosystem services values of Mekelle city region.

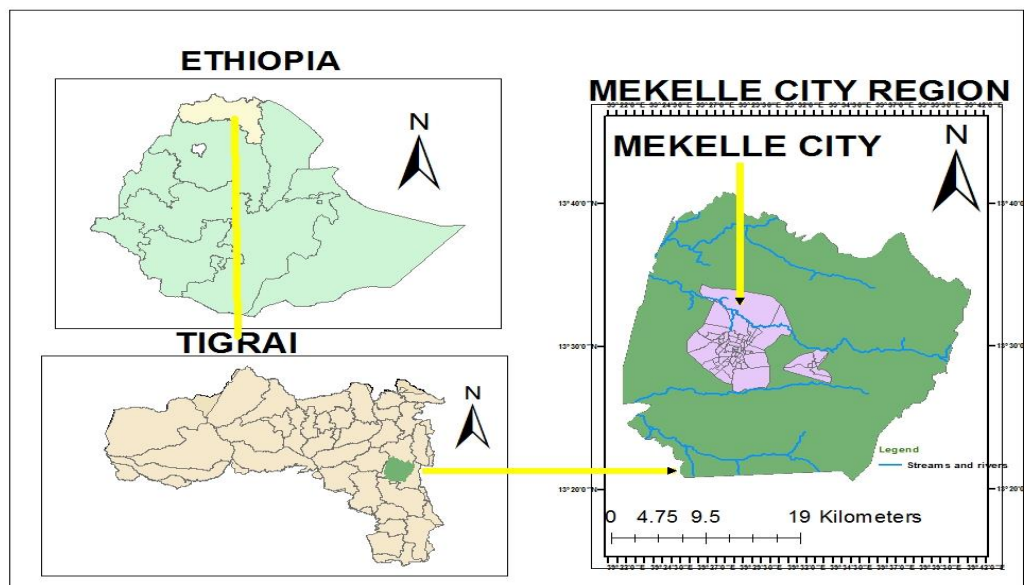
## II. MATERIAL AND METHODS

### 2.1 Study area

Mekelle city and its hinterlands which are significant watershed ecosystem services for the metropolitan Mekelle city are named as city region (figure 1). The landscape is located within Tigray region, the northern part of Ethiopia found in west 39.362942, East 39.687048, North 13.680920 and south 13.342621 about 760.61 km north of Addis Ababa and the area covered in this investigation is 897.12 square kilometres (89,712 hectares). The study area is pigeon-holed by varied topographic conditions. The elevation ranges from

1700 meters in the Geba river to 2685 meters Ellala and Gebat river catchment. The core city of the study area Mekelle city is 2062 m above sea level. The climate is largely semi-arid with asymmetrical rainfall and recurrent drought periods. The mean annual rainfall is estimated to be less than 532 mm (Veen, 2014). Climatically, the area has a semi-arid climate with little variation and it is known for its environmental vulnerability (Abreha, 2014). The agro-climatic zone of the study area is a mild climatic condition. The monthly mean minimum temperature is 15°C and the maximum monthly temperature may go as high as 28°C. The study area population is 556127 (TCSA, 2019).

**Figure 1: Location Map of Mekelle city region**



Source: Author, 2019

### 2.2 Data sources

#### 2.2.1 For the scenario simulation

The transition probability of transformation was represented by the Markov chain. The existence of a dynamic system represents the probability of a land-use type that will change from one class to another or stay in the same, depending on the class at the early time. In this analysis, the transition probability was used to estimate the chance of land use dynamics within a specified period. Driver-based simulating land dynamics was made in Mekelle city region. To model future, land-use changes additional spatial explicit datasets were created, which represent both the anthropogenic and biophysical drivers. This study used past land use/land cover changes observed data obtained from the Landsat series for the year 2019 and the simulated LULC dataset that was created using CLUMondo. This CLUMondo simulation tool is available as a free and open-source model. CLUMondo simulates changes in land systems in response to an exogenous demand, land system, and a series of biophysical and anthropogenic variables. CLUMondo is a dynamic, spatially explicit land use and land cover change model (Van Vliet &

Verburg, 2018). In addition to this, the Map Comparison Kit tool was used that includes numerous algorithms for comparing raster maps, which are particularly useful for land use dynamics simulation models.

Biophysical drivers such as mean monthly temperature and mean annual rainfall was prepared using IDW 3D analysis in Arc GIS, and anthropogenic drivers such as population density i.e Population density per pixel at 100-meter resolution was used which is found at (<https://energydata.info/dataset/ethiopia-population-density-2015>) by (Tatem, 2015)) drivers' data were collected and organized for those corresponding years for analysis from various sources. The elevation and slope obtained from digital terrain model (DEM, 30 m) from USGS (<https://earthexplorer.usgs.gov/>) SRTM1 Arc-second global. The slope was extracted using spatial analyst tool and elevation using create tin (3D analyst) in Arc GIS. To prepare distance to roads Euclidean tool was applied.

To analyze and assess them in a model set-up, all drivers were converted to raster format. The average values of temperature and precipitation from was converted to

raster data using statistical tools in Arcgis10.5 software, the raster spatial resolution was 30 m accordant with the remote sensing data. All raster data were changed to the same projection coordinate system UTM WGS1984\_UTM\_Zone\_37N and a spatial resolution of 30 m using ArcGIS. In that way, to avoid the modifiable aerial unit problem and for the sake of balanced analysis, data were resampled at the uniform spatial level and were associated in the same format, having the same projection and resolution parameters as those of the LULC raster maps.

**2.2.2 For Ecosystem services valuation**

LULC data for the year 2019 is the key data used for ecosystem service valuations. Then, the ecosystem services value per unit area by biome was multiplied by the total area of each biome and lastly summed over entirely ecosystem services. The classified images were likened with the representative biomes of LULC types to obtain their corresponding ESV coefficients identified.

**Table 1:** The LULC types with their area in hectares for 2019 and biome counterparts with the equivalent value coefficients ( US \$ ha<sup>-1</sup> year<sup>-1</sup>) adapted conservative value coefficients mainly based on the TEEB database by (Kindu, Schneider, Teketay, & Knoke, 2015).

S. N	Class name	Equivalent biome	(US \$ ha <sup>-1</sup> year <sup>-1</sup> )
1	Built-up area	Settlement	0
2	Natural forest	Tropical forest	986.69
3	Plantation forest	Tropical forest	986.69
4	Water body	Lakes/rivers fresh water	8103.5
5	Cultivated land	Cropland	225.56
6	Grass land	Grass/rangelands	293.25
7	Bushes and Shrubs	Wood land	986.69
8	River side vegetation	Wood land	986.69
9	Bare land	Desert	0

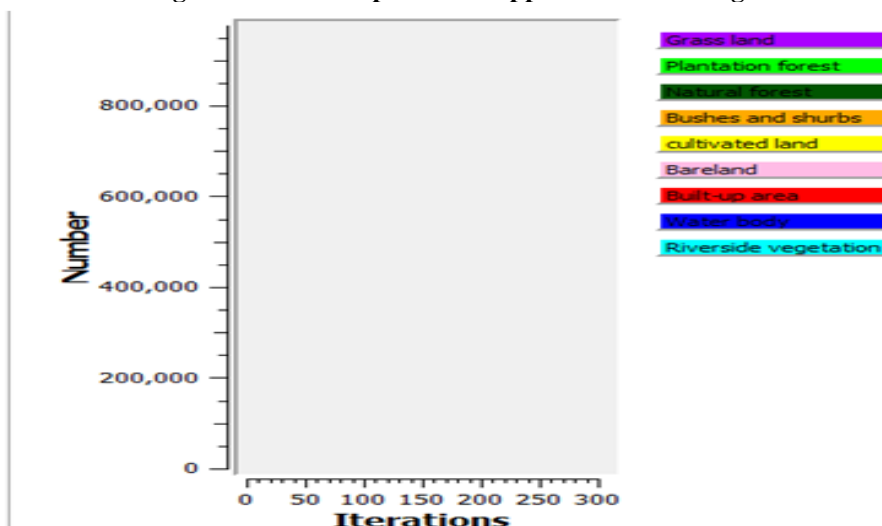
The study employed modified conservative value coefficients developed by (Kindu et al., 2015) to contextualize it to the Ethiopian and Mekelle city-region context. Based on the model quantifications of ESVs changes induced by LULC using a conservative coefficient which is a benefit transfer approach was applied. This value coefficient is assumed to be suitable for Mekelle city region. Resulting the ESVs methodology used by (Costanza et al., 1997; Kindu et al., 2015; J. Li, Chen, Zhang, & Pan, 2019; Tolessa, Gessese, Tolera, & Kidane, 2018; Yang, 2008).To estimate ESV the following equation is used.

$$ESV = \sum (A_k \times VC_k)$$

The ESV is the total assessed ecosystem service value, Ak is the area in (ha) and VCk the corresponding value coefficient (US\$ ha<sup>-1</sup> yr<sup>-1</sup>) for each LULC class k. The overall value of ecosystem services for the year 2019 and 2030 was attained by; the area of each LULC type in hectare was multiplied by its corresponding value coefficients to calculate the total ecosystem service value for a particular LULC type. The values for the LULC types in each reference year were summed to estimate the total ESV of the landscape year for each reference year. This study estimated future ESV using the predicted map of the year 2030 LULC charges under different scenarios.

**2.3 Data analysis**

**Figure 2: Iteration procedure applied for scenario generation**



In the interior each yearly time step, land systems were allocated in an iterative procedure in which land systems are allocated according to the transition potential at the time (t) and location (i) for each land system (LS), and the demands for goods and services for that specific year. The CLUMondo involves four wide-ranging inputs, namely land demand, location characteristics, spatial policies and restriction and land-use conversion scenery (Verborg, Vliet, Malek, & Ornetsmuller, 2015).

### Estimating land use demand

Mekelle city region has been facing huge land use transformation in the last five decades and the change. The forthcoming land-use dynamics scenarios for the year 2030 were simulated the future land uses that would meet the different level of conservation objectives of ecosystem service values. The land-use demands were calculated outside the CLUMondo model using an external model, which calculates the allocation of land use dynamics depending on the study area and the scenario. The climate, including mean monthly temperature and annual rainfall, is assumed to maintain the current state.

**Scenario A- business as usual (BAU):** this scenario assumes land-use dynamics of the past 2019 will also continue, in which the LUC of 2019 was used to project 2030 land cover without any planning restrictions and is assumed that land use dynamics will continue until 2030. The scrutiny of this study on land use dynamics and its paraphernalia on ecosystem service values of the past confirm a decline in ecosystem service value. Henceforward, there is likely deterioration in ecosystem service value, due to urbanization and generally, the built-up area is projected to increase significantly in 2030. In contrast, plantation forest, bushes and shrubs, natural forest, water bodies, grassland, riverside vegetation, bare land and cultivated land are projected to decrease sharply. If this scenario persists, it will have a severe impact on ecosystem service values.

**Scenario B- implementation of PES schemes:** the current trend of LULC changes would be halted with a hypothetical WTP and implementation of PES schemes. Under the PES scenario, LULC changes are more moderate and slower, and plantation forest, bushes and shrubs, waterbody, grassland, riverside vegetation will increase significantly. And it is assumed built-up areas will increase and cultivated land and bare lands will decrease. This scenario sets balanced emphasis on enhancing the value of ecosystem services and urbanization and built-up area expansion through regional spatial planning. This scenario assumed an increase in population and economic development. Altogether, the demand for built-up areas, including urban and rural residential land, construction land, and transport facility areas, would expand rapidly. To meet the growing population's demand for food, cultivated land will show less of a decline than the BAU scenario.

**Scenario C -Spatial planning:** This scenario puts balanced importance on continuing urbanization and enhancing ecosystem service values through integrated land uses plans.

This scenario assumes regional spatial planning will be implemented and ecosystem service values will increase. This scenario aims to increase overall ESVs from the current situation by 2030. For instance, currently, steep slopes that were under cultivation could be used for plantation forest, and degraded riverside vegetation by managing the land properly and made intensive plantations. In addition, this scenario anticipates effective protection of the remaining ecosystem in all existing and proposed protected areas. This scenario assumes that a more regulated urban development will be implemented. The scenario can be viewed as a harmonious development scenario that is characterized by condensed and slower built-up area and urbanization expansion in which the environment will be considered. This scenario gives priority to the protection of ecosystem restricted from being converted to other land uses. Furthermore, the area of built-up would show a slight decrease and natural forest, plantation forest, water bodies and riverside vegetation would increase more than other scenarios. This scenario will reduce the speed of the negative effects of urban expansion on ecosystem service values.

### Defining location characteristics

The location features were determined by assuming that the locations are devoted to land use type with the highest suitability as the location characteristics, such as DEM, distance to roads, mean annual rainfall, mean monthly temperature, population density and slope which are directly related to a specific location. The momentous drivers recognized for each land use category were used as input for projecting in terms of their level of significance on the respective land-use class. According to their correlation coefficient, the set of drivers were inserted in the regression equation as the independent X variables to estimate the predicted Y (land use/dependent variable). For 2030, the base year (T1) of prediction was 2019. Hereafter regression analysis was carried out on the driver's datasets. The multinomial Logistic regression predicts the probability of events under the comprehensive influence of multiple factors. The target variable of this study is the LULC changes in 2019 which is a binary variable. The dependent variable signifies the presence and absence of LULC.

The relationship between futuristic land use /land cover changes was established through multinomial logistic regression for each LULC type and a set of defined driving. The study used this model to calculate the probability with the biophysical and anthropogenic location characteristics as follows:

$$\text{Log} \left[ \frac{P_i}{1 - P_i} \right] = \beta_0 + \beta_1 X_{1,i} + \beta_2 X_{2,i} \dots \dots + \beta_n X_{n,i}$$

Where  $P_i$  is the probability of a grid cell  $i$  for the incidence of the land use type.  $X$ 's are the location factors. The coefficients ( $\beta$ ) are regression coefficients and were estimated through multinomial logistic regression using the actual land-use changes as the dependent variable. The model was to determine the observed relationships between

a binary dependent LULC changes classes and several independent continuous variables. For the dependent represents LULC changes results, Y has a binary value of 1 and 0.

**Table 2: List of variables included in the model**

Variable	Description	Name of variable
<b>Dependent</b>		
Y	0 no LULC changes ,1 LULC changes	Dichotomous
<b>Independent</b>		
X1	Population density	continuous
X2	DEM	continuous
X3	Distance to roads	continuous
X4	Slope	continuous
X5	Mean annual rainfall	continuous
X6	Mean monthly temperature	continuous

The independent variables data were named as population density (POP), DEM(DEM), distance to roads (DIS), slope (SLO), mean annual rainfall (RAI) and mean monthly temperature (Temp). Those factors impact the distribution of human-dominated land cover. Hence, to investigate the association between a categorical dependent LULC change variable and a set of independent driver’s multinomial logistic regression model was used. The dependent variable has only two values of 0 and 1.

**Land use alteration rate**

The CLUMondo modelling tool has many parameters and these were specified preceding running the simulation. The parameter setting is dependent on the assumptions made for a specific scenario. The translation resistance is the

settings specific to a particular land use type, which determines the temporal dynamics of the simulation. The conversion matrix indicates the land-use conversions that were allowed between the land uses. A value of 1 indicates that the conversion is allowed, and 0 indicates no conversion. The future land use dynamics was predicted as per the procedural requirements of CLUMondo, which requires information on locational characteristics, spatial policies and restrictions, land use conversion rates. Some locational characteristics, such as rainfall and temperature may vary over time and policy may change in future but this is hard to undertake for the distant future. The urbanization is concerned it is likely that there will be rapid change. The weight of the neighbourhood was set for different land-use types ranging from 0-1. Finally, the ecosystem service values of the future were calculated based on the predicted areal coverage of each land-use type by the year 2030.

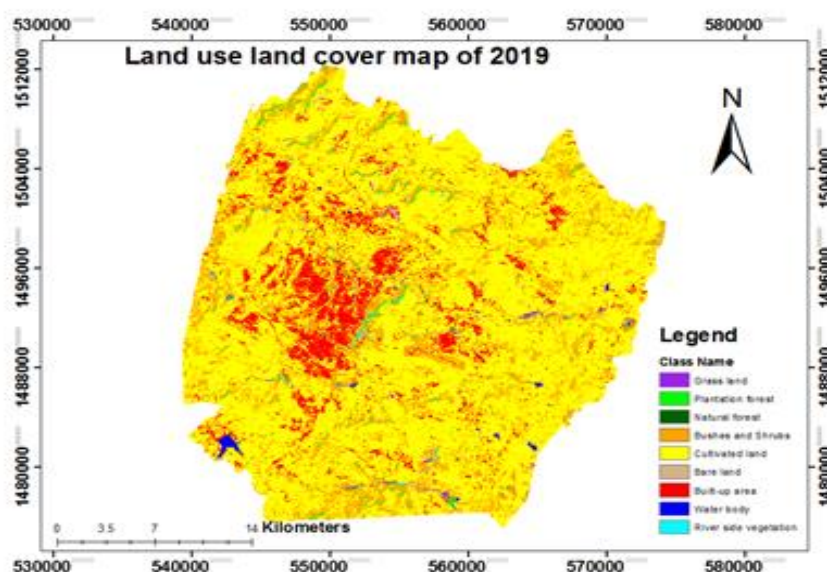
**Major steps used for predictions**

The study used the following methodology to predict LULC changes.

**1. Temporal analysis**

The land use/land cover changes were analysed using ERDAS imagine software. The CLUMondo software was used to simulate the spatial-temporal evolution of complex phenomena based on the transformation rules. Then the Markov model calculated between two periods the amount of change using the initial state and the transition probability and used for prediction of land use land cover change. Based on the initial(T1) and base year(T0), future land use was simulated for 2030(T2) which is aligned with the 2030 agenda for Sustainable development goals.

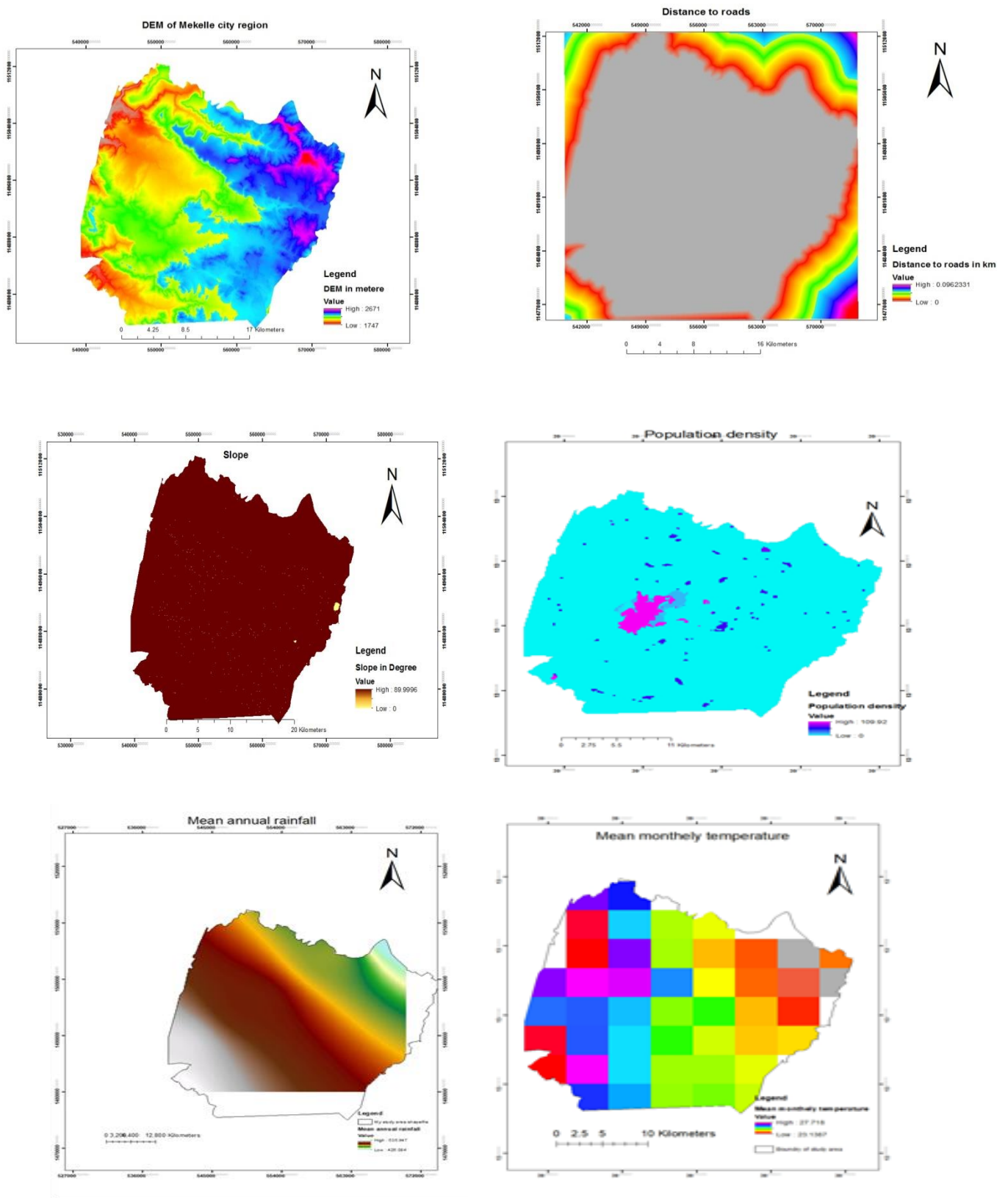
**Figure 3: Land use land cover classified image of the baseline year 2019**



## 2. Groundwork of database

Datasets for land use/land cover vagaries and drivers were organized in a comparable format convenient of simulation.

Figure 4: Drivers of future land us dynamics



### 3. Model validation and accuracy

The presence of a significant relationship between the drivers and response was assessed using the Relative operating characteristic plot (ROC) method. To assess how well the multinomial logistic regression model fits the dataset used here ROC was employed. It was used way to assess the accuracy of the model predictions by plotting sensitivity versus (1-specificity). The full area under a given ROC curve formulated an important statistic that represented the probability of the prediction that will be in the correct order when a test variable was observed for one

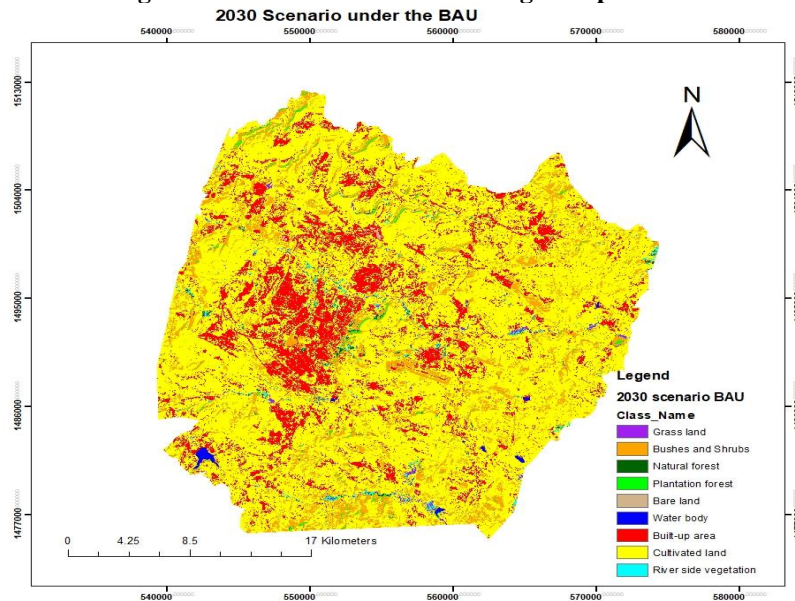
subject randomly selected drivers from the case group.

## III. RESULTS

### 3.1 Projected effects of land use dynamics on ecosystem services values

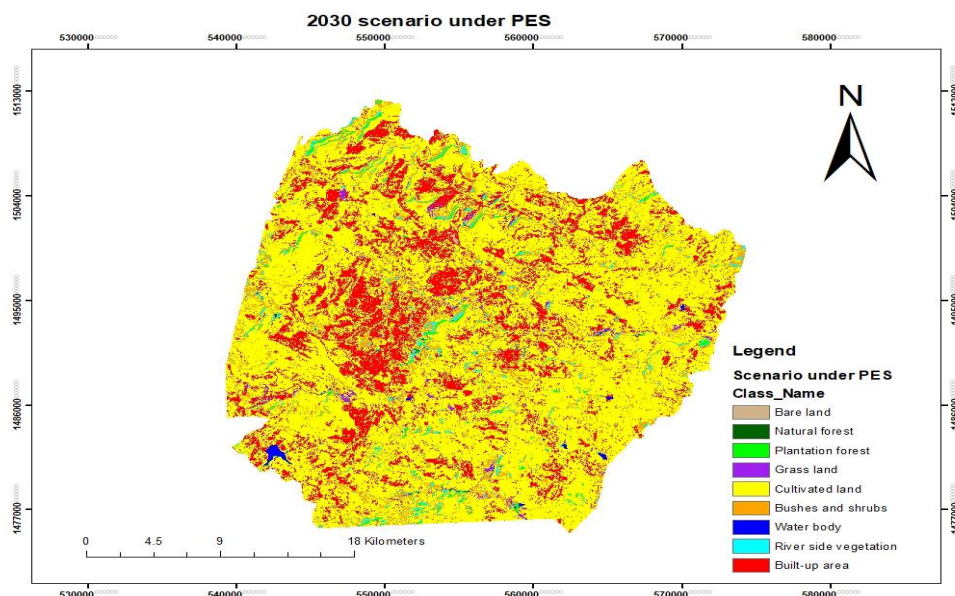
The land demand for different scenarios, land transfer rules and driving forces, simulation of land-use dynamics for the year 2030 were made. The predicted ecosystem service values of land use/land cover under different scenarios are presented in the figures below.

**Figure 5: Land-use/land cover changes maps of BAU**



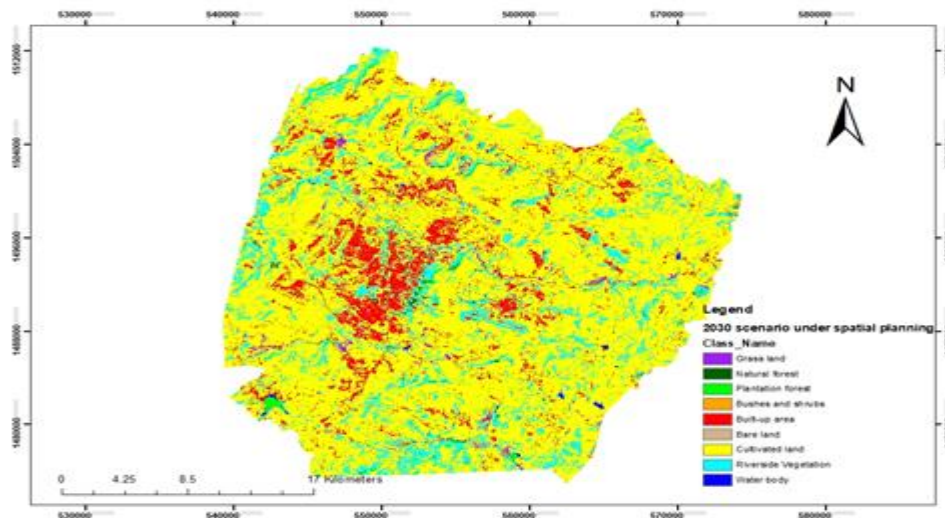
The results of scenario A (business as usual) showed that by the year 2030 it is predicted a dramatic increase in built-up areas, rising from 8897 hectares (9.92%) in 2019 to 14692.59 hectares (16.37%) in 2030 at the expense of natural forest, plantation forest, water bodies, cultivated land, bushes and shrubs, and riverside vegetation. The variations in land use under this scenario showed a negative effect on the ecosystem compared to the base year of 2019. This situation may be attributed to the expansion of built-up areas.

**Figure 6: Land-use/land cover maps of the scenario under PES**



In scenario B (PES schemes), the built-up areas will continue to increase due to rapid urbanization. It will have an increase of 1439.77 hectares (9.8%) from the 2030 BAU scenario. Despite the expansion of built-up areas under scenario B, the natural forest will increase by 15 hectares (14.86%), plantation forest by 326.52 hectares (67.14%), water bodies by 301.62 hectares (140.93%), grassland by 498.06(114%), riverside vegetation by 1012.32 hectares (149.87%). However, under this scenario, cultivated land, bushes and shrubs, and bare land will decline by 508.59 hectares (0.82%),3025.85(27.38%),58.86(46.42%) respectively. The deviations in land use under this scenario showed a positive impact on the ecosystem compared to the BAU scenario.

**Figure 7: Land-use/land cover maps of the scenario under spatial planning scenario**  
 2030 scenario under spatial planning



In scenario 3 (Spatial planning), the forest cover will increase considerably by 2030. This is for the reason that, in addition to the existing ecosystem at present, the scenario aims to raise forest cover to increase ecosystem services values. Compared to Scenario 2 under this scenario, various LULC types such as natural forest will increase by 88 hectares (75.86%), plantation forest by 743.86 hectares (91.5%), water bodies by 4.05 hectare (2.72%), and riverside vegetation by 8771.92 hectares (519.73%). These changes in land use under this scenario showed a positive impact on the ecosystem by 2030 as compared to the PES scenario. Conversely, this scenario predicts a decrease in built-up areas, cultivated land grassland, and bushes and shrubs.

**3.2 Markov prediction for the year 2030**

Table 3: Land-use /land cover changes, from the baseline year 2019 to 2030, under the BAU, PES, and spatial planning scenarios

Classes	2019		2030 BAU		2030 PES schemes		2030 spatial planning	
	Ha	%	Ha	%	Ha	%	Ha	%
Built-up area	8897	9.92	14692.59	16.37	16132.36	17.9	8099.46	9.12
Natural forest	106	0.13	101	0.11	116.01	0.13	204.01	0.23
Plantation forest	497	0.55	486.36	0.54	812.88	0.91	1556.74	1.74
Water bodies	263	0.29	214.02	0.24	515.64	0.612	529.69	0.59
Cultivated land	64843	72.27	61929	69	61420.41	68.5	60952.62	67.9
Grassland	330	0.37	436.77	0.49	934.83	1.05	881.82	0.98
Bushes and shrubs	13215	14.73	11050	12.3	8024.15	8.94	6951.47	7.75
Riverside vegetation	1235	1.38	675.45	0.81	1687.77	1.88	10459.69	11.6
Bare land	326	0.36	126.81	0.14	67.95	0.078	76.5	0.09
Total	89712	100	89712	100	89712	100	89712	100



From the above prediction, the PES scenario presents the greatest built-up area expansion (17.9%) of the total study area which is much higher than those predicted under the BAU (16.37%) and spatial planning scenarios 9.12%. The importance of spatial planning and PES schemes is clear when comparing the land-use predictions for BAU scenarios. For example, under the BAU scenario, the increase in built-up areas occurs at the expense of all other land-use types such as natural forest, plantation forest, water bodies, cultivated land, bushes and shrubs, and riverside vegetation. But under the PES scenario, the expansion of the built-up area at the expense of cultivated land, bare lands, and bushes and shrubs. The spatial planning scenario shows a decrease in built-up areas expansion and while an increase in riverside vegetation at the expense of built-up areas, cultivated land, bushes and shrubs, and grasslands. The result shows that some of the LULC classes under spatial planning scenarios will change significantly.

**3.3 Assessment of future Ecosystem Service Values for the year 2030**

**Table 4: The ecosystem service values per hectare of different land-use types for the year 2030**

S. N	Class name	Area in hectares *corresponding coefficient values in millions USD							
		2019		2030					
		ESV	%	BAU		PES Schemes		Spatial Planning	
ESV	%			ESV	%	ESV	%	ESV	%
1	Built-up area	0	0	0	0	0	0	0	0
2	Natural forest	0.105	0.33	0.100	0.36	0.114	0.4	0.2	0.54
3	Plantation forest	0.490	1.53	0.479	1.72	0.800	2.78	1.5	4
4	Water body	2.1	6.56	1.7	6.2	4.17	14.5	4.3	11.6
5	Cultivated land	14.6	45.6	13.9	49.8	13.85	48.1	13.75	37
6	Grass land	0.097	0.30	0.128	0.47	0.270	0.94	0.250	0.67
7	Bushes and Shrubs	13.4	41.9	10.9	39	7.9	27.4	6.85	18.4
8	River side vegetation	1.21	3.78	0.666	2.4	1.7	5.9	10.32	27.79
9	Bare land	0	0	0	0	0	0	0	0
	<b>Total</b>	<b>32</b>	<b>100</b>	<b>27.98</b>	<b>100</b>	<b>28.8</b>	<b>100</b>	<b>37.2</b>	<b>100</b>

From the above table, the estimation of future ecosystem service values under the BAU scenario shows a decrease in total ecosystem service values from the baseline year 2019 by 4.02 million USD (12.18%). However, PES schemes have a positive difference of 825,770.7 USD (+2.95%) from the BAU scenario. The prediction results for spatial planning show the increase of ESVs by 5.2 million USD (16.25%) from the baseline year 2019. This study revealed that the major contributor to these future changes is a decrease in built-up area expansion and increases in riverside vegetation, natural forest, plantation forest, and water bodies.

The result in the table below discovered that there will be diverse effects of LULC variations on the various ecosystem service values; it will increase some ecosystem service values while others will probably decline. In the business as usual scenario, there will be decline in all ecosystem service value in USD comparing with 2019 - water supply by 0.69, food production (0.15), raw material (0.06), genetic resources (0.05), water regulation (1.79), water treatment (0.33), erosion control (0.33), climatic regulation(0.29),biological control(0.02),gas regulation (0.02),disturbance regulation(0.01), nutrient cycling (0.25),pollination(0.02),soil formation (0.01),habitat (0.02),recreation(0.04) and culture(0.02).

In contrast from the BAU scenario, PES schemes is a policy intervention scenario and total ecosystem service value will be increased by 0.82 USD(2.93%).In spatial planning scenario, compared to the baseline year 2019 the individual ecosystem service values in USD will increase water supply by 0.89,food production(0.2),raw material(0.08),genetic resources(0.07), water regulation (2.27), water treatment(0.42), erosion control(0.042), climatic regulation(0.37), biological control(0.01), gas regulation (0.02), disturbance regulation(no change), nutrientcycling(0.3), pollination(0.02), soilformation(0.02), habitat(0.03), recreation(0.03) and culture by (0.03) with an overall ecosystem service values increases of 5.2 USD(16.25%) compared to the base year 2019.

**3.4 Assessed services of individual ecosystem functions and projected changes**

Table 5: Estimated value of ecosystem functions (ESV<sub>f</sub> in US\$ million US\$ year<sup>-1</sup>) under each service category for the year 2019 and 2030

Ecosystem services		The annual estimated value of ecosystem functions (ESV <sub>f</sub> in US\$ million US\$ year <sup>-1</sup> )							
		ESV <sub>f2030</sub>							
		2019		BAU		PES Schemes		Spatial planning	
		ESVs	% coverage	ESVs	% coverage	ESVs	% coverage	ESVs	% coverage
<b>Provisioning Services</b>	Water supply	5.5	17.15	4.81	17.19	4.95	17.20	6.39	17.17
	Food production	1.2	3.74	1.05	3.75	1.08	3.75	1.4	3.76
	Raw material	0.5	1.55	0.44	1.57	0.45	1.58	0.58	1.56
	Genetic resources	0.4	1.24	0.35	1.25	0.36	1.25	0.47	1.26
<b>Regulating Services</b>	Water regulation	14	43.72	12.21	43.64	12.57	43.65	16.27	43.71
	Water treatment	2.6	8.12	2.27	8.11	2.34	8.13	3.02	8.12
	Erosion control	2.6	8.12	2.27	8.11	2.34	8.13	3.02	8.12
	Climatic regulation	2.3	7.18	2.01	7.18	2.07	7.19	2.67	7.18
	Biological control	0.12	0.38	0.10	0.36	0.10	0.36	0.13	0.35
	Gas regulation	0.16	0.5	0.14	0.5	0.14	0.49	0.18	0.48
	Disturbance regulation	0.06	0.19	0.05	0.18	0.05	0.17	0.06	0.16
<b>Supporting services</b>	Nutrient cycling	1.9	5.93	1.65	5.9	1.7	5.9	2.2	5.91
	Pollination	0.17	0.53	0.15	0.54	0.15	0.52	0.19	0.51
	Soil formation	0.1	0.31	0.09	0.32	0.09	0.3	0.12	0.32
	Habitat	0.18	0.56	0.16	0.57	0.16	0.55	0.21	0.56
<b>Cultural services</b>	Recreation	0.22	0.69	0.18	0.64	0.19	0.66	0.25	0.67
	Cultural	0.03	0.09	0.05	0.19	0.05	0.17	0.06	0.16
	<b>Total</b>	<b>32</b>	<b>100</b>	<b>27.98</b>	<b>100</b>	<b>28.8</b>	<b>100</b>	<b>37.2</b>	<b>100</b>

**3.5 Multinomial logistic regression results**

The multinomial logistic regression was formed to investigate the association between LULC change and the driving forces that play an important role in LULC change. It is used to predict the behaviour of a dependent variable concerning an independent variable. The model was run for each of the land use/land cover types and all the selected drivers of changes. The prediction analysis using the regression model identified the important drivers for each land-use class using the level of significance for each LULC class dynamics.

**Table 6: Correlation coefficient of independent driving forces variables**

	X1	X2	X3	X4	X5	X6
X1	1	0.002	0.000	0.001	0.003	0.403
X2	0.002	1	0.005	0.011	0.000	0.004
X3	0.000	0.005	1	0.000	0.000	0.048
X4	0.001	0.011	0.000	1	0.000	0.033
X5	0.003	0.000	0.000	0.000	1	0.014
X6	0.403	0.004	0.048	0.033	0.014	1

The regression adjustment checked the existence of collinearity among independent driving forces variables and was done using a Pearson correlation coefficient in SPSS 24. As shown in the above table the collinearity is not strong.

**Table 7: Multinomial Logistic regression results between LUCC and driving force**

Driving forces	Built-up area			Natural forest			Plantation forest			Waterbody			Cultivated land		
	$\beta$	S.E.	Sig.	$\beta$	S.E.	Sig.	$\beta$	S.E.	Sig.	$\beta$	S.E.	Sig.	$\beta$	S.E.	Sig.
Pop	0.001	0.000	0.012	0.350	0.211	0.062	0.251	0.313	0.072	0.191	0.124	0.123	0.001	0.000	0.014
DEM	-0.020	0.001	0.055	0.005	0.001	0.000	0.405	0.001	0.000	0.001	0.000	0.025	0.004	0.001	0.001
DIS	0.001	0.000	0.100	0.043	0.017	0.181	0.032	0.026	0.261	0.001	0.000	0.000	0.000	0.000	0.033
SLO	0.013	0.040	0.746	0.061	0.021	0.005	0.061	0.031	0.005	0.035	0.02	0.072	0.052	0.034	0.057
RAI	0.011	0.009	0.203	0.033	0.005	0.154	0.024	0.005	0.166	0.003	0.003	0.158	0.003	0.006	0.568
Temp	-0.350	0.667	0.490	0.113	0.215	0.000	0.233	0.315	0.000	0.132	0.247	0.594	0.537	0.456	0.262

Driving forces	Grassland			Bushes and shrubs			Riverside vegetation			Bare land		
	$\beta$	S.E.	Sig.	$\beta$	S.E.	Sig.	$\beta$	S.E.	Sig.	$\beta$	S.E.	Sig.
Pop	0.186	0.059	0.001	0.560	0.329	0.091	0.450	0.221	0.072	0.249	0.143	0.015
DEM	0.002	0.000	0.000	0.006	0.001	0.000	0.005	0.001	0.000	0.001	0.000	0.020
DIS	0.004	0.013	0.732	0.024	0.018	0.192	0.033	0.017	0.173	0.005	0.014	0.687
SLO	0.011	0.012	0.34	0.051	0.018	0.005	0.051	0.011	0.005	0.092	0.016	0.000
RAI	0.003	0.002	0.184	0.004	0.004	0.156	0.013	0.005	0.245	0.006	0.002	0.008
Temp	0.022	0.187	0.904	0.110	0.215	0.000	0.11	0.314	0.000	0.005	0.178	0.975

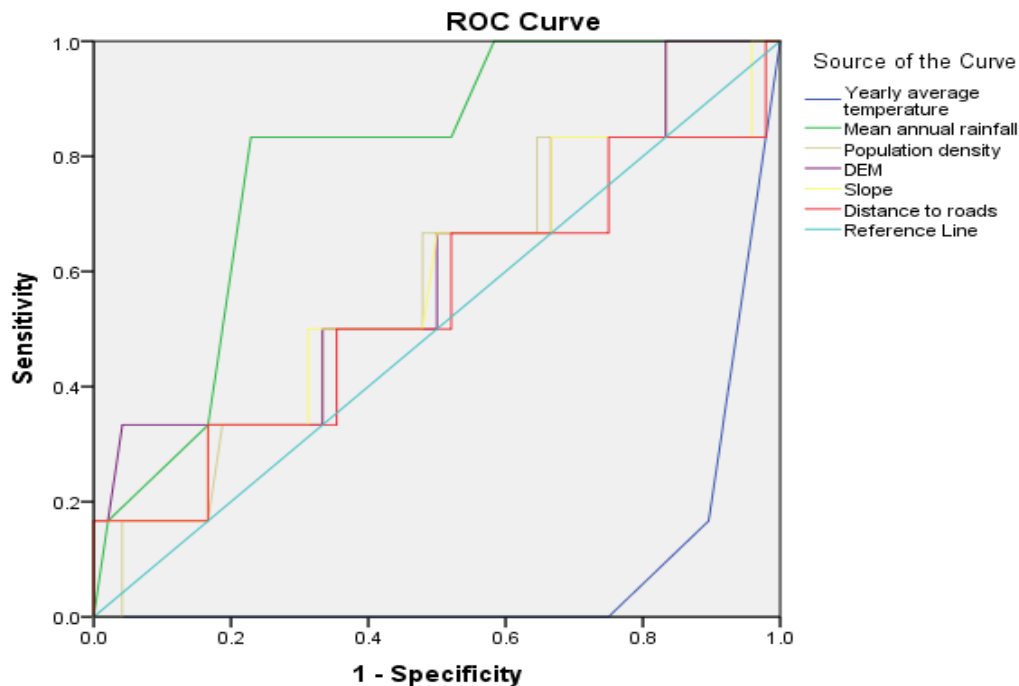
From the above result,  $\beta$  signifies the regression coefficients between land use/land cover types and its drivers,  $\text{Exp}(\beta)$  represents the probability of land use/land cover type change. All driving forces are significant at  $p < 0.05$ . When  $\text{Exp}(\beta) > 1$ , this indicates the probability of increases of land use types upon an increase in the value of the driving factors, and when  $\text{Exp}(\beta) < 1$ , this indicates the probability of land use/land cover type reductions.

**3.6 Calibration and validation of the model**

The ROC plot used to testify the logistic regression the one simulated for the year 2030 compared to the actual 2019 LULC map. The ROC assessed the pair of maps agree with the location of points being converted from one land use to

another type of LULC. The bigger values of the test result variable indicate stronger evidence for a positive actual state. The test result variable shows Population density has at least one tie between the positive actual state group and the negative actual state group. The positive actual state is the Built-up area. An ideal model would have an area of 1.000, and this logistic model has the largest area under the curve of 0.73 with a standard error of 0.38 in the figure below. The test result variables namely yearly average temperature, mean annual rainfall, population density, DEM, Slope has at least one tie between the positive real state group and the negative actual state group. Statistics may be biased.

Figure 8: ROC plot



Diagonal segments are produced by ties.

The area under the ROC curve was used to quantify the overall model performance. It was interpreted as the average value of sensitivity for all possible values of specificity 1. The closer AUC is to 1, the better the overall analytical performance of the test, and a test with an AUC value of 1 is perfectly accurate. The practical lower limit for the AUC of a diagnostic test is 0.5. The line segment from 0, 0 to 1, 1 has an area of 0.5(Park, Goo, & Jo, 2004). The classifiers that gave curves closer to the top-left corner in the above indicate better performance. The above model interpretation indicates the ROC curve (the green line) in the top left corner of the plot, which indicates that the model does a good job of predicting mean annual rainfall will have a significant impact on future LULC.

Table 8: Area Under the Curve (AUC)

Test Outcome Variable(s)	Area
Yearly average temperature	0.073
Mean annual rainfall	0.792
Population density	0.582
DEM	0.606
Slope	0.568
Distance to roads	0.538

The ROC test of five driving forces of LULC changes is greater than 0.538 except yearly average temperature which shows clearly that the spatial distribution for all land use categories can be explained by the related driving factors. Under the three scenarios, future urbanization will be concentrated in the core city of the study area in Mekelle city, where the urbanization will continue. However, the use of different LULC scenarios decreases the uncertainties

associated with the model and enhances the applicability of simulation results. To reduce uncertainty logistic regression was also used. The accuracy of the model was tested by comparing the 2019 LULC map with its simulated complement. Generally, the validation shows a satisfactory result, which shows that the simulation model can be used to simulate future LULC.

### 3.7 The nexus between simulation of land use dynamics and ecosystem services evaluation with sustainable development goals

To ensure the sustainability of Mekelle city region considering the three pillars of sustainability is crucial. The economic, environmental, and social pillars are vital in understanding sustainability. If any one of the three pillars is failed then the ecosystem services provided as a whole can become unsustainable. This study considered the economic, ecological, and social values in valuing ecosystem services. The results of ecosystem service valuation in Mekelle city region show an unsustainable path, and there is a need to get on a sustainable one. In this study, the goal of the ecosystem service valuation is to contribute to a more sustainable and equitable natural resource use. The final goal of ecosystem services valuation is to improve the well-being of every individual living now in the study area, currently, and in the future. This study followed the strong sustainability view i.e the economy is merely a sub-system of society, which is itself a subsystem of the environment, all being interdependent.

During valuation, this study respected the principles of sustainable development. The development in Mekelle city region should meet the needs of current residents without compromising the ability of future generations to meet their

own needs. Economic and social well-being cannot be improved with measures that destroy the environment.

The ecological valuation and ecosystem assessment was the first step during ES valuation. The social and monetary values depend on the ES production. The production functions described here how ecosystems generate services were acknowledged. The economic valuation of ES is considered for value in a broader sense including the social and ecological qualities embedded in these services, which are perceived by those who benefit from ES at different scales. The social valuation of ES involves the collective decisions through deliberation allow actors to decide who gives up what as part of a negotiation and collective learning process. The value formation occurred through expert discussion. The valuation process was transparent to be used for decision-making. When valuing ES, the questions of governance cannot be ignored. Valuation exercises always take place in a given institutional setting. Because ecosystem services are common goods, this institutional setting ideally favours social rationality and communicative action, warranting that a societal perspective was considered.

In social-ecological systems theory social refers to the human dimension, including economic, political, technological, cultural, and ecological, and are embedded parts of the biosphere and shape it. Societies will continue to be shaped by the biosphere and there is a growing recognition of the need for development approaches that take account of the interdependencies between human well-being and the biosphere e.g., SDGs goals. The results of the study are in line with sustainable development goals 11 and 15.

Goal 11: Make cities and human settlements inclusive, safe, resilient, and sustainable and 11. support positive economic, social and environmental links between urban, peri-urban, and rural areas by strengthening national and regional development planning. The concept of city region considers both urban and rural areas it is with the goals of this sustainable development goal. Besides, goal 15: Protect, restore and promote sustainable use of terrestrial ecosystems, sustainably manage forests, combat desertification, and halt and reverse land degradation and halt biodiversity loss

By 2020, Integrate ecosystem and biodiversity values into national and local planning, development processes, poverty reduction strategies, and accounts. This study applied ecosystem services valuation to understand the environmental degradation and sustaining ecosystem services provision. Based on the WCED of 1987 report, this study applied the three pillars of sustainability namely economy, society, and environment.

### 3.8 Discussion

CLUMondo model was used to predict the land use dynamics until 2030 and the year for sustainable development goals achievement target was selected. To guide future development in Mekelle city and its hinterlands prediction of LULC and its effects on ecosystem service

values is vital. In the study area, there is an economic boom in and this could have short-term and long-term effects on ecosystem service values. When major changes in ecosystems and ecosystem services are at stake with the expected change in, as a consequence of land-use change or economic/environmental policy, structured and transparent valuations are appropriate, in the proper temporal and spatial contexts (Robert Costanza & Steve Farber f, 2017).

This study weighs the approaches and different software packages for modelling LUC such as CLUMondo, GeoSOS, LanduseSim, LuccME, SAGA GIS, SLEUTH GA, QGIS MOLUSCE, Dinamica Ego, UrbanCA, future land use simulation model (FLUS), and IDRISI Selva 17.02. This learning reviews the potentials and the limits of these modelling software packages. First, a revision of the methods and tools available for each model was performed, taking into account how the models carry out the different procedures involved in the modelling process: quantity of change estimate, change potential evaluation, spatial allocation of change, reproduction of temporal and spatial patterns, model evaluation and advanced modelling options. Supplementary considerations, such as flexibility, user-friendliness, were also taken into account. Then, the models were applied to the study to illustrate the previous descriptions with a typical LUC scenario that consists of processes of change. The outputs were compared to assess the change estimates, the change potential, and the simulated maps. However, due to its compatibility, the study utilized CLUMondo version 1.4.0. The simulation tool was tested in simulating future land-use changes and ecosystem services in northern Thailand by (Arunyawat & Shrestha, 2018).

To understand the change in ecosystem service value resulting from a change in land use and land cover, this study begins with a scenario that exhibits different types of land-use change. The scenario generator creates different patterns of conversion according to the LULC inputs. The CLUMondo creates maps of land-use change according to user-assigned probabilities that certain transitions will occur. The changes in land use/land cover were made in response to demands for various goods and services provided by the land, such as cultivated land converted for providing built-up areas. Changes in these demands can result in land use/land cover conversion, an increase in cultivated land demand led to conversion from forests to cultivated land and a growing population led to an increase in the built-up area. The future land-use scenarios for the year 2030 were simulated as business-as-usual, application of possible PES schemes, and spatial planning. Under the BAU scenario if the existing condition is not improved in terms of land use planning and introducing new policies such as PES schemes ecosystem service values will probably drop sharply.

Different studies employed different driving criteria to study LULC scenario modelling. For example, predicting urban growth using a statistical modelling approach by (Manu, Twumasi, Lu, & Coleman, 2015) and Integration of logistic regression, Markov chain and cellular automata models to simulate urban expansion (Arsanjani, Helbich,

Kainz, & Bolorani, 2013). But this study is unique in the selection of drivers. This study applied LULC modelling including the statistical, spatial, and driver-based method with quantitative future land use prediction.

Previous studies used different modelling tools and logistic regression for predictive land-use modelling. This study will be amongst the first to simulate future LULC changes under different scenarios at the city-region level. The developed scenarios in Mekelle city region are different from the existing ones. logistic regression is one of the most common approaches used in land use allocation models to determine the relationship between land use dynamics and its driving forces. However, these models have a drawback in that they do not determine/allocate land use based on the direct multi-class problem of land-use change and its driving factors. The model has a weakness in that it does not determine land use based on the direct relationship between land-use change and its driving factors. Accordingly, a multinomial logistic regression method was introduced to the study to address this defect. The application of this method in future land-use change studies may improve the accuracy of predicting the effects of land use and cover change on ESVs.

This study showed that spatial analysis and multinomial logistic regression produced a meaningful model of drivers of LULC changes. The population of Mekelle city region will grow over 1.4 million inhabitants by 2030, and the land demands will exceed the sustainable capacity of land use. Spatial planning has got little attention in the study area. Continuous urbanization and population growth and the impact of climate changes will place additional stress on the ecosystem unless appropriate integrated planning and management actions are put in place. Urban planners need scenario estimation tools to achieve sustainable development.

#### IV. CONCLUSION AND RECOMMENDATION

The study aimed to forecast conceivable future land use scenarios through mapping and quantification of ESVs. This study explored how future land-use changes would affect ecosystem service values, under three different scenarios for the year 2030. The land-use scenario in 2030 is predicted, and the value of ecosystem services is estimated. CLUDMondo model has been used to predict future land use and land cover maps of the year 2030 based on the data. Based on logistic regression analysis, the contribution of driving forces to the land use and cover change was obtained. To validate land-use dynamics simulation ROC method was used for the model's prediction accuracy.

This study directed the reductions in the total and individual ecosystem service values in response to LULC changes under the BAU scenario. The results show that the value of ecosystem services will continue to decrease. It was found that the significant expansion of built-up area and lessening of various ecosystem services are expected in 2030 that negatively affects their values. Under the PES and spatial planning, scenarios show the largest increase in

ESVs. The PES is relatively better compared with the BAU scenarios, due to new policy implementation on ecological land. As a result, the spatial planning scenario shows the highest amount of ESVs. Based on these findings, it is proposed new land-use strategies, to achieve sustainable use of ecosystem services in the study area. This study can guide future land-use planning and raise public awareness of the need for ecosystem protection. The identification of trends in land cover change and ecosystem service defined in this study could help develop future scenarios to tackle future challenges in ecosystem service supply. The results deliver useful inputs to the LULC planners and resources managers to adopt eco-friendly policies and regulations and for encouraging sustainable development goals.

In this respect, it is suggested (1) a comprehensive understanding of future land use dynamics scenarios and their environmental impacts is needed by spatial planners to make sound decisions in practising sustainable land management (2) governmental agencies should use could scenario simulation to regulate how to manage lands, to provide a desirable range of benefits to people, or to help design permitting and mitigation programs that sustain nature's benefits to society (3) efficient land-use systems through regional spatial planning need to be implemented and (4) there is the need for policy changes, and the findings of the study can contribute towards achieving SDGs. The study finding can also serve as the starting point in developing sustainable land uses for environmental conservation in Ethiopia's major urban centres. The outcomes offer a quantitative systematic report that policymakers and land managers can use to identify and prioritize the best practices to sustain ecosystem services.

#### ACKNOWLEDGEMENTS

My thanks go to Tigray Bureau of trade, industry and urban development, Tigray (Ethiopia), the bureau of civil service and Ethiopian civil service university for allowing me to pursue this study and for the financial support, without which this achievement would not have been possible. I would like to thank all institutions and individuals at the national, regional, district and city levels who provided us with information for this study.

#### Declaration of interest

The authors declare that they have no competing interests.

#### Ethics approval

Not applicable.

#### Consent to participate

The author needs to accept and confirm the manuscript have read and agreed to its content and are accountable for all aspects of the accuracy and integrity of the manuscript. The article is original, has not already been published in a journal, and is not currently under consideration by another journal.

**Consent for publication**

The author is a volunteer for the publication of original research findings entitled " **Scenario Simulation of ecosystem services value changes: implication for sustainability development goals** ". The author grants the **International Journal of Innovative Science and Research Technology (IJISRT)** a license to publish the article and identify itself as the original publisher.

**Availability of data and material**

The dataset(s) supporting the conclusions of this article is (are) available in Google Scholar, JSTOR, Google that all publicly available datasets are fully referenced in the reference list.

**Authors' contributions**

- 1) The Author (shishay kiros) makes substantial contributions to conception and design, and/or acquisition of data, and/or analysis and interpretation of data;
- 2) The Co-author (Ranavijai Singh) contribution is revising critically for important intellectual content of the article and edited the whole manuscript.
- 3) The Co-author (kumelachew Yeshitela) contribution is revising critically for important intellectual content and edited the manuscript.

**Authors' information**

1. Kumelachew Yeshitela (PhD) is an associate professor of Urban Ecology at the Ethiopian Institute of Architecture, Building Construction and City Development, Addis Ababa University, Addis Ababa, Ethiopia. He is Chair of Ecosystem Planning and Management. He has published 47 scientific works in international reputable journals.
2. Ranavijai Singh, Ph.D, Associate Professor, College of Urban Development and Engineering, Ethiopian Civil Service University, Ababa, Ethiopia.
3. Shishay is a Ph.D. Candidate in Environmental sustainability and Climate change resilience at Ethiopian Civil Service University, Addis Ababa.

**REFERENCES**

- [1]. Abreha, A. G. (2014). *Hydrogeochemical and Water Quality Investigation on Irrigation and Drinking Water Supplies in the Mekelle Region, Northern Ethiopia* (MSC), University of Twente 2014.
- [2]. Ardavan Zarandiana, H. B., Nigel E. Storkc, Matthew A. Lingd, Ahmad R. Yavarie, Hamid R. Jafarif, Hamid Amirnejad. (2017). Modelling ecosystem services informs spatial planning in lands adjacent to the Sarvelat and Javaherdasht protected area in northern Iran. *Land use policy*, 487–500.
- [3]. Arsanjani, J. J., Helbich, M., Kainz, W., & Boloorani, A. D. (2013). Integration of logistic regression, Markov chain and cellular automata models to simulate urban expansion. *International Journal of Applied Earth Observation and Geoinformation*, 21, 265-275.
- [4]. Arunyawat, S., & Shrestha, R. P. (2018). Simulating future land use and ecosystem services , Northern Thailand. *Journal of Land Use Science*, 13(1-2), 146-165.
- [5]. Costanza, R., d'Arge, R., De Groot, R., Farber, S., Grasso, M., Hannon, B., . . . Paruelo, J. (1997). The value of the world ecosystem services and natural capital. *nature*, 387(6630), 253.
- [6]. Fu, Q., Hou, Y., Wang, B., Bi, X., Li, B., & Zhang, X. (2018). Scenario analysis of ecosystem service changes and interactions in a mountain-oasis-desert system: a case study in Altay Prefecture, China. *Scientific reports*, 8(1), 1-13.
- [7]. Houet, T., Verburg, P. H., & Loveland, T. R. (2010). Monitoring and modelling landscape dynamics. *Landscape ecology*, 25(2), 163-167.
- [8]. Hu, S., Chen, L., Li, L., Zhang, T., Yuan, L., Cheng, L., . . . Wen, M. (2020). Simulation of land use change and ecosystem service value dynamics under ecological constraints in Anhui Province, China. *International journal of environmental research and public health*, 17(12), 4228.
- [9]. Kindu, M., Schneider, T., Teketay, D., & Knoke, T. (2015). Drivers of land use/land cover changes in Munessa-Shashemene landscape of the south-central highlands of Ethiopia. *Environmental monitoring and assessment*, 187(7), 452.
- [10]. Li, J., Chen, H., Zhang, C., & Pan, T. (2019). Variations in ecosystem service value in response to land use/land cover changes in Central Asia over 1995-2035: PeerJ Preprints.
- [11]. Li, Z., Cheng, X., & Han, H. (2020). Future Impacts of Land Use Change on Ecosystem Services under Different Scenarios in the Ecological Conservation Area, Beijing, China. *Forests*, 11(5), 584.
- [12]. Manu, A., Twumasi, Y. A., Lu, K. S., & Coleman, T. L. (2015). Predicting urban growth of a developing country city using a statistical modeling approach. *International Journal Of Geomatics And Geosciences*, 5(4), 603-613.
- [13]. Olosutean, H. (2015). Methods for modeling ecosystem services: a review. *Management of Sustainable Development*, 7(1), 5-12.
- [14]. Pandeya, B., Buytaert, W., Zulkafli, Z., Karpouzoglou, T., Mao, F., & Hannah, D. (2016). A comparative analysis of ecosystem services valuation approaches for application at the local scale and in data scarce regions. *Ecosystem Services*, 22, 250-259.
- [15]. Park, S. H., Goo, J. M., & Jo, C.-H. (2004). Receiver operating characteristic (ROC) curve: practical review for radiologists. *Korean journal of radiology*, 5(1), 11.
- [16]. Robert Costanza , R. d. G., Leon Braat , Ida Kubiszewski , Lorenzo Fioramonti , Paul Sutton , & Steve Farber f, M. G. g. (2017). The Twenty years of ecosystem services: How far have we come and how far do we still need to go? *Elsevier*.
- [17]. Tatem, A. (2015). Demographics of Ethiopia. from World bank group

- [18]. TCSA. (2019). *Tigray population data* Demography. Tigray stastical agency.
- [19]. Tolessa, T., Gessese, H., Tolera, M., & Kidane, M. (2018). Changes in ecosystem service values in response to changes in landscape composition in the central highlands of Ethiopia. *Environmental Processes*, 5(3), 483-501.
- [20]. Van Vliet, J., & Verburg, P. (2018). A short presentation of CLUMondo *Geomatic approaches for modeling land change scenarios* (pp. 485-492): Springer.
- [21]. Veen, T. G. a. A. v. d. (2014). The Effect of Enclosures in Rehabilitating Degraded Vegetation: A Case of Enderta District, Northern Ethiopia *Forest Research*.
- [22]. Verburg, P., Vliet, J., Malek, Z., & Ornetsmuller, C. (2015). The CLUMondo land use change model: Manual and exercises. *Institute for Environmental Studies, University Amsterdam*.
- [23]. Yang, H. (2008). An ecosystem service value assessment of land-use change on Poyang Lake Basin under 3S technology, China. *The International Archives of the Photogrammetry, Remote Sensing, and Spatial Information Sciences*, 37.
- [24]. Yirsaw, E., Wu, W., Shi, X., Temesgen, H., & Bekele, B. (2017). Land use/land cover change modeling and the prediction of subsequent changes in ecosystem service values in a coastal area of China, the Su-Xi-Chang Region. *Sustainability*, 9(7), 1204.